

1 **Tracking marine litter with a global ocean model:**
2 **Where does it go? Where does it come from?**

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Abstract

Plastic is the most abundant type of marine litter and their presence in the environment is a major concern because they remain in the oceans for a long time. It is found in all of the world's oceans and seas, even in remote areas far from human activities. To address questions that are of tremendous interest to the international community as it seeks to attend to the major sources of marine plastics in the ocean, we use particle tracking simulations to simulate the motions of mismanaged plastic waste and provide a quantitative global estimate of 1) where does the marine litter released into the ocean by a given country go and 2) where does the marine litter found on the coastline of a given country come from. The overall distribution of the modeled marine litter is in good agreement with the limited observations that we have at our disposal. The tables summarizing the statistics for all world countries are accessible from the supplemental information in .pdf or .csv formats.

23 **1. Introduction**

24 Plastic is the most abundant type of marine litter and their presence in the environment is a
25 major concern because they remain in the oceans for a long time, affecting marine life and
26 threatening human health. Steady growth in the amount of discarded solid waste combined with
27 the slow degradation rate of many waste items are gradually increasing the amount of marine litter
28 found at sea, on the seafloor, and along the coastal shores. This distribution and accumulation have
29 become an economic, environmental, human health, and aesthetic problem that presents a complex
30 and multi-dimensional challenge. Marine litter results from human behavior, whether accidental
31 or intentional. According to the United Nations Environment Programme (UNEP), eighty percent
32 of the marine litter originates from land sources including waste released from dumpsites near the
33 coast or river banks, the littering of beaches, tourism and recreational use on the coasts, fishing
34 industry activities, and ship-breaking yards. The primary sea-based sources include abandoned,
35 lost, or discarded fishing gear, shipping activities, as well as legal and illegal dumping. In the
36 global context, understanding of marine litter as a persistent and growing problem has become
37 clear. The United Nations Environment Assembly (UNEA) has recognized marine litter as one of
38 its top priorities through three resolutions (from UNEA-1 in 2014, UNEA-2 in 2016, and UNEA-
39 3 in 2017), specifically calling for actions to combat marine litter
40 (<https://environmentassembly.unenvironment.org/>).

41 Litter is found in all of the world's oceans and seas, even in remote areas far from human
42 activities. Thus, tracking the movement of plastic litter in the ocean is crucial. Ocean currents
43 control the distribution and accumulation of floating marine debris, but observational data are very
44 sparse and it is difficult to analyze and predict the movement of debris (van Sebille et al., 2020).
45 Factors that determine the transport and fate of debris include its size and buoyancy. Marine

46 plastics are classified as either macro, micro, or nano. A macro plastic is the largest of the three
47 classifications and consists of plastic that can be easily seen with the naked eye. Examples include
48 plastic bags, water bottles, and fishing nets. The next classification of plastic is micro plastics,
49 which are generally considered to be one to five millimeters in length. Primarily we see micro
50 plastics in the form of plastic pellets, which are the building blocks of plastic. Secondary micro
51 plastics are formed as macro plastics break down from exposure to sunlight, temperature, wave
52 and salt. Micro plastics can easily be incorporated into the food chain, and, because of this, have
53 become the main focus of environmental conversation. Finally, nano plastics are a byproduct of
54 micro plastics as they degrade. They can be as small as 1 μm (micrometer) and, due to their
55 extremely small size, it is possible for Nano plastics to enter the food chain. While there is much
56 to be learned about nano plastics, it is known that they pose a significant threat to the environment
57 and humankind (see special issue in Nature Nanotechnology, April 2019).

58 In 2016, the global production of plastics was approximately 330 million metric tons (Mt;
59 Plastics Europe, 2017) and that amount is estimated to double within the next 20 years (Lebreton
60 and Andrady, 2019). Plastics are usually divided in three categories: plastics in use, post-consumer
61 managed plastic waste, and mismanaged plastic waste (Geyer et al., 2017). Mismanaged plastic
62 waste (MPW) is defined as plastic material littered, ill-disposed, or from uncontrolled landfills.
63 Plastic debris enters the sea from the coastal environment through runoff, winds, and gravity
64 (Jambeck et al., 2015) and via rivers (Lebreton et al., 2017). There are however few direct
65 measurements of plastic entering the ocean and one has to rely on conceptual frameworks
66 (Jambeck et al., 2015; Lebreton et al., 2017; Schmidt et al., 2017; Lebreton and Andrady, 2019) to
67 compute, from the best available data, an order-of-magnitude estimate of the amount of MPW

68 entering the world ocean. This lack of data on waste generation, characterization, collection, and
69 disposal, especially outside of urban centers, leads to uncertainties (Jambeck et al., 2015).

70 Most of our understanding on the motion of floating marine debris comes from numerical
71 simulations (Hardesty et al., 2017; van Sebille et al., 2020). Given the scarcity of observational
72 data, numerical models can be used to simulate the motions of debris and test scenarios. In this
73 paper, we use particle tracking simulations to address questions that are of tremendous interest to
74 the United Nations and the international community as they seek to track, identify, and eventually
75 attend to the major sources of marine plastics in the ocean. The questions we address in this paper
76 are:

- 77 a. Where does MPW released into the ocean by a given country go?
- 78 b. Where does MPW found on the coastline of a given country come from?

79 The layout of this paper is as follows: In section 2, we describe the numerical model. The
80 uncertainties associated with the marine litter sources are introduced in section 3. The seeding
81 strategy, wind effects, parameterized unresolved processes, and decay scenarios are discussed in
82 section 4. The results are presented in section 5. The last section provides a summary and discusses
83 the limitations of the current model.

84 **2. Model description**

85 The global framework we use to track marine litter is OceanParcels v2.1.5, which can create
86 customizable particle tracking simulations using outputs from ocean circulation models.
87 OceanParcels v2.1.5 is a state-of-the-art Lagrangian ocean analysis tool designed to combine (1)
88 wide flexibility to model particles of different natures and (2) efficient implementation in
89 accordance with modern computing infrastructure. The latest version includes a set of interpolation

90 schemes that can read various types of discretized fields, from rectilinear to curvilinear grids in
91 the horizontal direction, from z- to s- levels in the vertical and different variable distributions such
92 as the Arakawa's A-, B- and C- grids (Delandmeter and van Sebille, 2019).

93 The ocean circulation model outputs used in OceanParcels are from the GOF3.1, a global
94 ocean reanalysis based on the HYbrid Coordinate Ocean Model (HYCOM) and the Navy Coupled
95 Ocean Data Assimilation (NCODA; Chassignet et al., 2009; Metzger et al., 2014). NCODA uses
96 a three-dimensional (3D) variational scheme and assimilates available satellite altimeter
97 observations, satellite, and in-situ sea surface temperature as well as in-situ vertical temperature
98 and salinity profiles from Expendable Bathythermographs (XBTs), Argo floats, and moored
99 buoys. Surface information is projected downward into the water column using Improved
100 Synthetic Ocean Profiles (Helber et al., 2013). The horizontal resolution and the frequency for the
101 GOF3.1 outputs are $1/12^\circ$ (8 km at the equator, 6 km at mid-latitudes) and 3-hourly, respectively.
102 For details on the ocean circulation model validation, the reader is referred to Metzger et al. (2017).

103 **3. Marine litter sources**

104 Plastic debris in the ocean is usually assumed to be from land-based sources, although some
105 studies have suggested that sea-based sources also play an important role (e.g., Bergmann et al,
106 2017; Lebreton et al., 2018). No matter the source, the primary challenge of modeling the global
107 displacement of marine litter are the large uncertainties associated with the amount and location
108 of mismanaged plastic waste (MPW) entering the ocean. In this paper, we consider only the land-
109 based sources. In order to derive meaningful information from the numerical simulation and
110 address the above questions, one needs to be able to seed the model with plastic waste entering the
111 ocean that are representative of each country and have been computed in a consistent manner
112 globally. At the present time, there are four studies that can provide a first order estimate of the

113 current global plastic waste input from land into the ocean: Jambeck et al. (2015), Lebreton et al.
114 (2017), Schmidt et al. (2017), and Lebreton and Andrady (2019). However, these studies all differ
115 in their estimates of MPW input into the ocean.

116 Starting with the earlier study by Jambeck et al. (2015) for the coastal environment, the authors
117 estimated an annual input of plastic to the ocean by taking into account 1) the mass of the waste
118 generated per capita annually, 2) the percentage of waste that is plastic, and 3) the percentage of
119 waste that is mismanaged and thus has the potential to enter the ocean. The calculation is based on
120 a 2010 World Bank dataset (Hoorweg and Bhada-Tata, 2012) on country-specific waste
121 generation and management. Jambeck et al. (2015) estimated that ~11% of the 2.5 billion metric
122 tons (t) total solid waste generated by the 6.4 billion people living in 192 coastal countries (i.e.,
123 275 million metric tons [Mt]) is plastic and, scaling by the population living within 50 km of the
124 coast, they calculated that 99.5 Mt of plastic waste was generated in the coastal regions.
125 Mismanaged waste is defined as material that is either littered or inadequately disposed of meaning
126 that it is not formally managed. This includes disposal in dumps or open, uncontrolled landfills,
127 where waste is not fully contained. Mismanaged waste can eventually enter the ocean via inland
128 waterways, wastewater outflows, and transport by wind or tides. Jambeck et al. (2015) estimated
129 that, in 2010, 31.9 Mt were mismanaged and that between 4.8 and 12.7 Mt (15–40%) made it to
130 the ocean (1.7–4.6% of the total plastic waste). Assuming no improvements to the waste
131 management infrastructure, the cumulative quantity of plastic waste available to enter the marine
132 environment from land was predicted to increase by an order of magnitude by 2025.

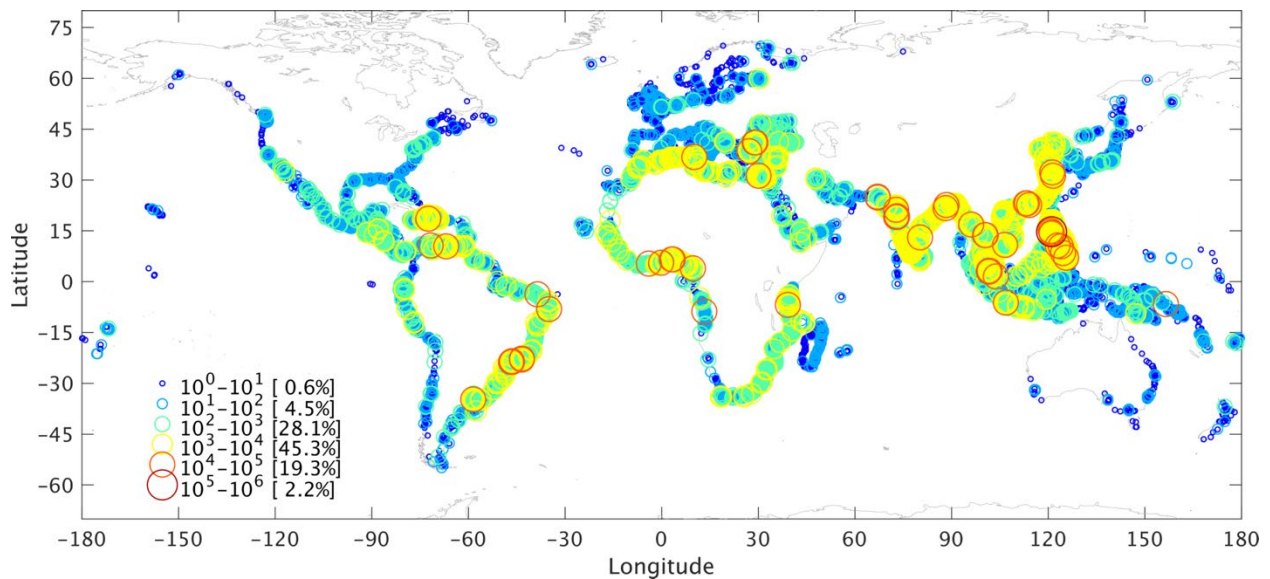
133 Plastics in the coastal areas usually enter the ocean via direct littering that is moved offshore
134 by the wind and/or tidal currents. But plastics can also enter via rivers. Lebreton et al. (2017)
135 estimated between 0.36 and 0.89 Mt per year enter via river transport in the coastal area (about 3-

136 19% of the total MPW 4.8-12.7 Mt of Jambeck et al. [2015]). In addition, they estimate at least
137 0.8 to 1.5 Mt per year reach the oceans from inland areas via rivers. Schmidt et al. (2017)
138 independently derived a total MPW carried in the global river system to the ocean of between .5
139 and 2.7 Mt, which supports the estimates of Lebreton et al. (2017), i.e., between 1.1 and 2.4 Mt.
140 The spatial distribution of the Schmidt et al. (2017) data is qualitatively similar to those of Jambeck
141 et al. (2015), but the fraction contributed by the larger rivers is considerably higher.

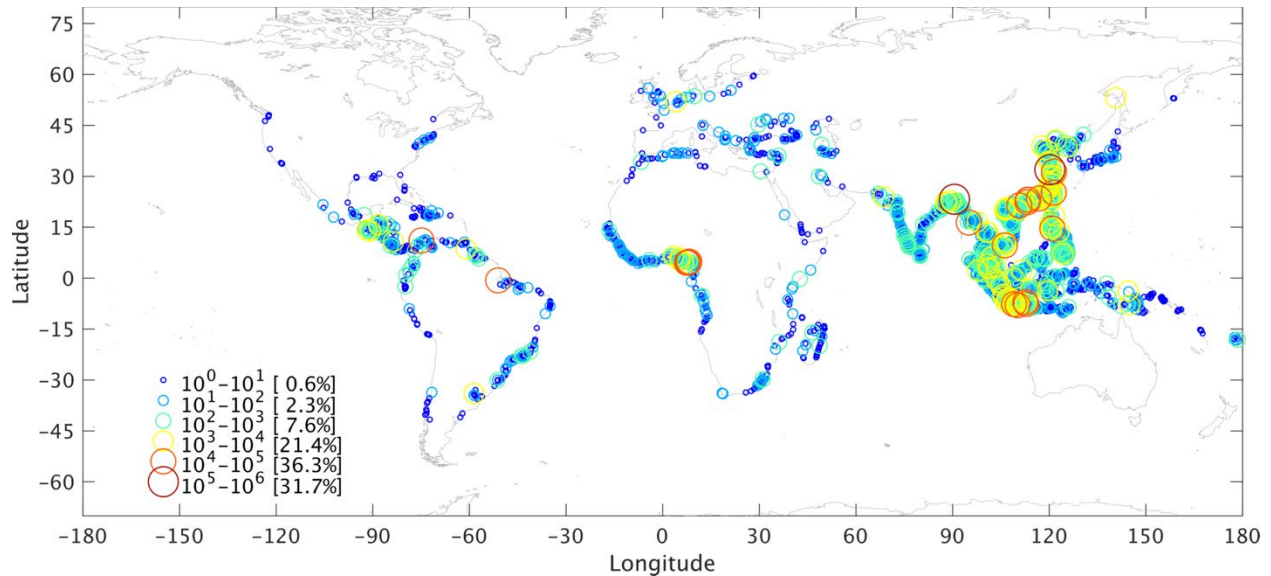
142 Finally, Lebreton and Andrady (2019) present projections of global MPW generation at ~1 km
143 resolution from 2015 to 2060. They estimate that between 60 and 99 Mt of MPW were produced
144 globally in 2015 (see the 2015 annual distribution of MPWs from coastal regions in Figure 1) and
145 that this figure could triple by 2060. One of the main motivations for that study was to quantify
146 the fraction of MPW generated in coastal areas against the fraction generated inland that may reach
147 the oceans via rivers (see river distribution in Figure 2). Following Jambeck et al.'s (2015)
148 framework and using their fine-resolution global distribution of MPW, Lebreton and Andrady
149 (2019) estimated a total of 20.5 Mt of MPW generated from the coastal population in 2010. This
150 value converted to an annual global input of MPW to the ocean from the coastal regions to be
151 between 3.1 and 8.2 Mt is slightly lower than the 4.8 to 12.7 Mt estimate of Jambeck et al. (2015).
152 However, as stated earlier, estimating MPW associated with the population within a fixed distance
153 from the coast (50 km, as in Jambeck et al. [2015]) does not take into account MPW generated
154 inland and transported by rivers. Given the fine granularity of their data, Lebreton and Andrady
155 (2019) were able to estimate that, for 2015, approximately 5% of MPW was discarded directly into
156 small watersheds near the coastline, that 4% was discarded in proximity of the coastline in medium
157 watersheds, and that the majority (91%) was discarded in large watersheds away from the

158 coastline. Therefore, as shown by Lebreton et al. (2017) and Schmidt et al. (2017), the large rivers
159 are a major source of plastic waste from inland to the ocean and should not be neglected.

160 In summary, while the four studies provide different estimates of MPW reaching the ocean,
161 they are consistent. As indicated by Schmidt et al. (2017), this is not too surprising because they
162 all start from the same waste database and use similar conceptual frameworks. There are, of course,
163 large uncertainties associated with the numbers provided by the above studies, but they provide a
164 globally consistent database that can be used to seed our model. For this study, we derived MPW
165 inputs for the model using Lebreton and Andrady (2019) for the coastal regions (within 50 km
166 from the coastline; Figure 1) and Lebreton et al. (2017) for the inland regions via rivers (Figure 2).
167 It is important to note that a large portion of the MPW (~40% according to Andrady [2011]) that
168 reach the ocean are denser than seawater and thus will sink to the ocean floor near the coast instead
169 of being carried away by the surface currents and/or wind.



170
171 **Figure 1.** Distribution of the annual MPW input from the coastal regions (50 km from the
172 coastline) based on Lebreton and Andrady (2019). Blue to red circles with increasing size represent
173 MPW of different order of weights (in tons/year) and their percentage values with respect to the
174 total (~5.1 Million tons/year).



175

176 **Figure 2.** Distribution of the annual MPW input from the inland region through rivers based on
 177 Lebreton et al. (2017). As in Figure 1, blue to red circles with increasing size represent MPW of
 178 different order of weights (in tons/year) and their percentage values with respect to the total (~1.4
 179 million tons/year).

180 **4. Seeding strategy, Stokes drift, wind drag, random walk, and decay scenarios**

181 As laid out in section 3, we divide the global MPW inputs in the world ocean into two
 182 categories: 1) direct input from coastal regions, defined as within 50 km of the coastline; and 2)
 183 indirect input from inland regions via rivers. For direct input (Figure 1), we use MPW computed
 184 from the global database on a 30 x 30 arc seconds grid of Lebreton and Andrady (2019). The
 185 MPWs from within 50 km from the coastline are summed (in ton/year) on a $1/4^\circ \times 1/4^\circ$ grid and,
 186 as in Jambeck et al. (2015), we assume that only 25% of MPW enters the world ocean. We neglect
 187 contributions that are less than 10 tons/year (~ 0.6% of the total direct MPW) and the number of
 188 particles that are released on each grid cell is as follows: one particle is released in each month for
 189 cells that have MPW in the 10-10² ton/year range, three particles each month for cells with 10-10²
 190 tons/year; 3² for 10³-10⁴ tons/year; 3³ for 10⁴-10⁵ tons/year; 3⁴ for 10⁵-10⁶ tons/year, etc. In total,
 191 we release 28,713 particles each month along the coastline representing the 5.1 million tons of
 192 MPW that enters the ocean per year. For the indirect input (Figure 2), we use the midpoint

193 estimates for the global river catchments assembled by Lebreton et al. (2017). As for the direct
194 MPW input, we neglect contributions by rivers that are less than 10 ton/year (~0.6% of total
195 indirect MPW). In total, we release 3,287 particles each month at the river mouth, representing the
196 1.4 million tons of MPW that enters the ocean per year. More than two-thirds of MPW (in terms
197 of weight) enters via 21 rivers, mostly from South and East Asia.

198 For a review of the physical oceanography associated with the transport of floating marine
199 plastics and of all the processes that affect transport, the reader is referred to van Sebille et al.
200 (2020). In short, the particles are moved around by ocean currents, surface wave induced Stokes
201 drift, and wind drag. As described in section 2, the ocean surface currents used in this study are
202 from GOF3.1, a global ocean forecast system (Chassignet et al., 2009; Metzger et al., 2014) based
203 on the HYbrid Coordinate Ocean Model (HYCOM) and the Navy Coupled Ocean Data
204 Assimilation (NCODA). The Ekman transport resulting from the atmospheric forcing (Fleet
205 Numerical Meteorology and Oceanography Center 3-hourly NAVY Global Environmental Model,
206 NAVGEM) is included in the ocean surface currents. All simulations include a small random walk
207 component representing unresolved turbulent motions in the ocean.

208 A full account of the Stokes drift, which is induced by surface gravity waves in the direction
209 of wave propagation (see review by van den Bremer and Breivik, 2018, for detail), would require
210 an accurate wave model. However, the wave-induced Stokes drift can be assumed to act in the
211 same direction of the wind (e.g., Kinsman, 1965; Kubota, 1994; Breivik et al., 2011) and the joint
212 effect of wind and wave can be expressed as a single wind drag coefficient. Pereiro et al. (2018),
213 using observed data from 23 drifters together with wind and ocean current data, suggested a wind
214 drag coefficient ranges from 0.5% to 1.2%. This is in agreement with Arduin et al. (2009), who
215 estimated the magnitude of the wave-induced contribution by the Stokes drift to be ~0.6%-1.3%

216 of the wind speed in the wind direction. The magnitude of the contribution does depend on the
217 buoyancy ratio of the plastic object and the sea water, i.e., lighter objects correspond to higher
218 coefficients (Chubarenko et al., 2016). Because detailed information on different types of MPW
219 entering the global ocean is not available, this study adopts a wind drag coefficient of 1%, as in
220 Kubota (1994).

221 One additional factor that needs to be taken into account when modeling MPW is the time it
222 takes for plastics to break down into smaller pieces under the combined actions of waves and
223 effects of sunlight. These micro or nano plastics end up either in suspension in the water column
224 (e.g., Kukulka et al., 2012), sinking to the bottom (e.g., Thompson et al., 2004; Woodall et al.,
225 2014; Barrett et al., 2020), ingested and entangled by marine organisms (e.g., Moore et al., 2001),
226 or decomposed (Kimukai et al., 2020). To account for those complex processes that ultimately
227 lead to the removal of the MPW from the sea surface (where the abundance of the MPW is
228 observed and the movement of MPW is simulated), we apply a simple, hypothetical exponential
229 decay function to the weight (mass)

$$230 \quad W(t) = W(0) * e^{-t/t_0},$$

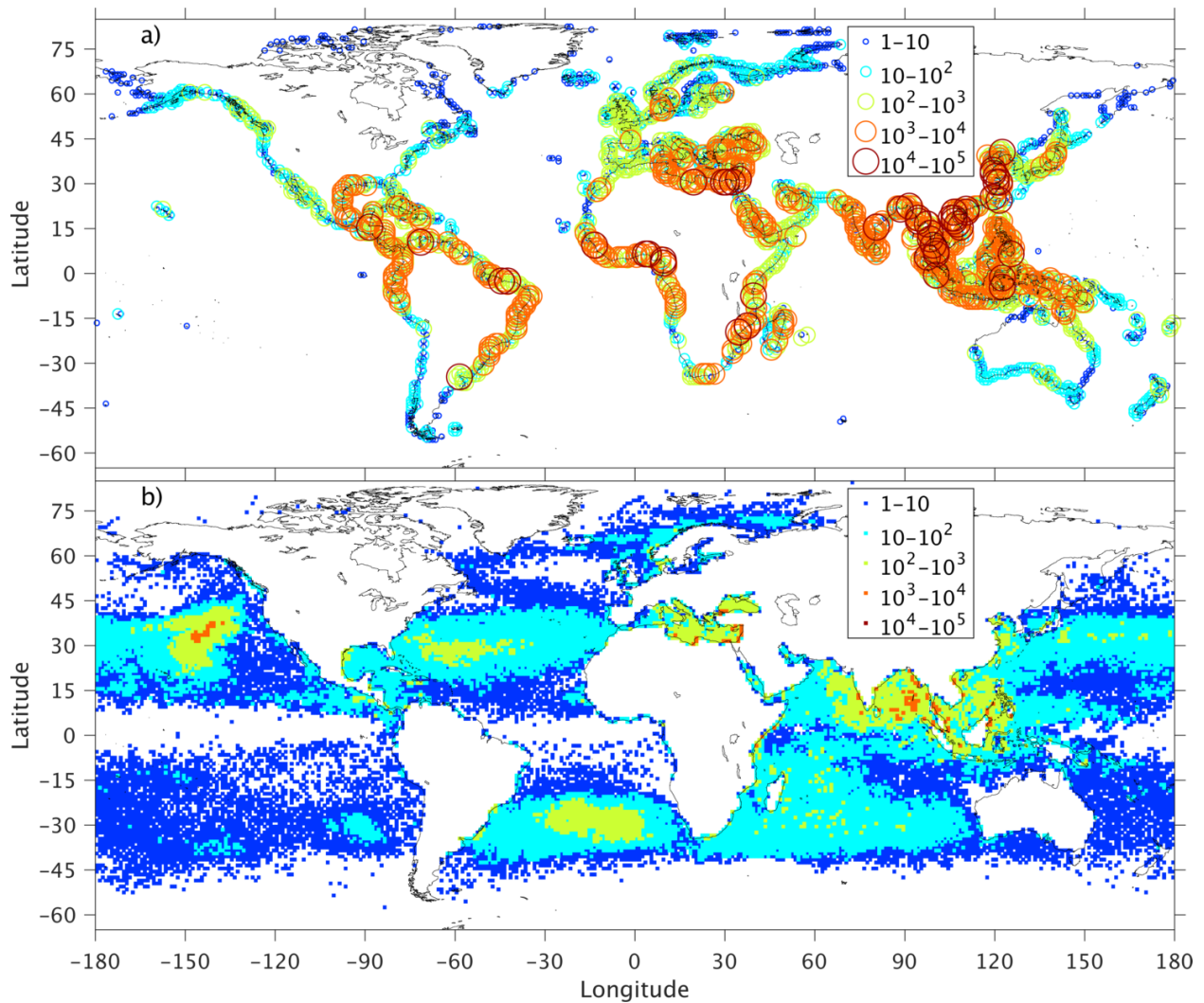
231 in which t is time (in years), $W(0)$ is the MPW weight released into the ocean, and t_0 represents
232 an e-folding time scale. After experimentation and comparison to observations (see discussion in
233 next section), we adopted an e-folding time scale of five years. This implies that 36.8% of the
234 MPW weight would remain at the surface after five years (13.5% after 10 years).

235 **5. Results**

236 Following the seeding strategy for MPW described in the previous section, we release 32,300
237 particles (28,713 for coastal inputs and 3,287 for inland inputs via rivers) every month from 2010
238 to 2019 along the global coastline. These particles represent a total of 3.9 Mt of lighter-than-water
239 MPW per year (3.0 Mt coastal and 0.9 Mt inland) released into the ocean. After release, using
240 OceanParcels v2.1.5 (see section 2), the particles are advected by the ocean currents (with a small
241 random walk component to account for unresolved turbulent motions) and wind (1%). All particles
242 are integrated from the release point to the end of 2019. The results presented in this section
243 correspond to a 10-year accumulation of MPW in the ocean. We first describe the MPW
244 concentration at sea and on the beach, and then provide statistics for each country on MPW
245 destinations and beached MPW sources.

246 **5.1 Partition between beached versus at sea MPW**

247 Here we consider a MPW particle to be “beached” if the sum of daily displacements in the last
248 30 days of the integration is less than a constant threshold distance. Figure 3 displays the number
249 of particles in $1^\circ \times 1^\circ$ grid boxes that are beached versus those that remain at sea at the end of the
250 2010-2019 accumulation period. Of the MPW released during 2010-2019 mass, 74.5% end up
251 beached (Figure 3a), whereas 25.5% of the MPW remains at sea (Figure 3b). This partition
252 between beached and at sea MPW is not very sensitive to the individual year, except for 2019,
253 during which the percentage of in-water MPW mass is higher (44%) because the integration is too
254 short for some of the MPW particles to reach to the shore. Thus, the estimate of $\sim 3/4$ of beached
255 MPW and $\sim 1/4$ remaining at sea is robust. Not surprisingly, we find that the wind is primarily
256 responsible for the beaching and that the impact of random walk/diffusion on beaching is quite
257 small.

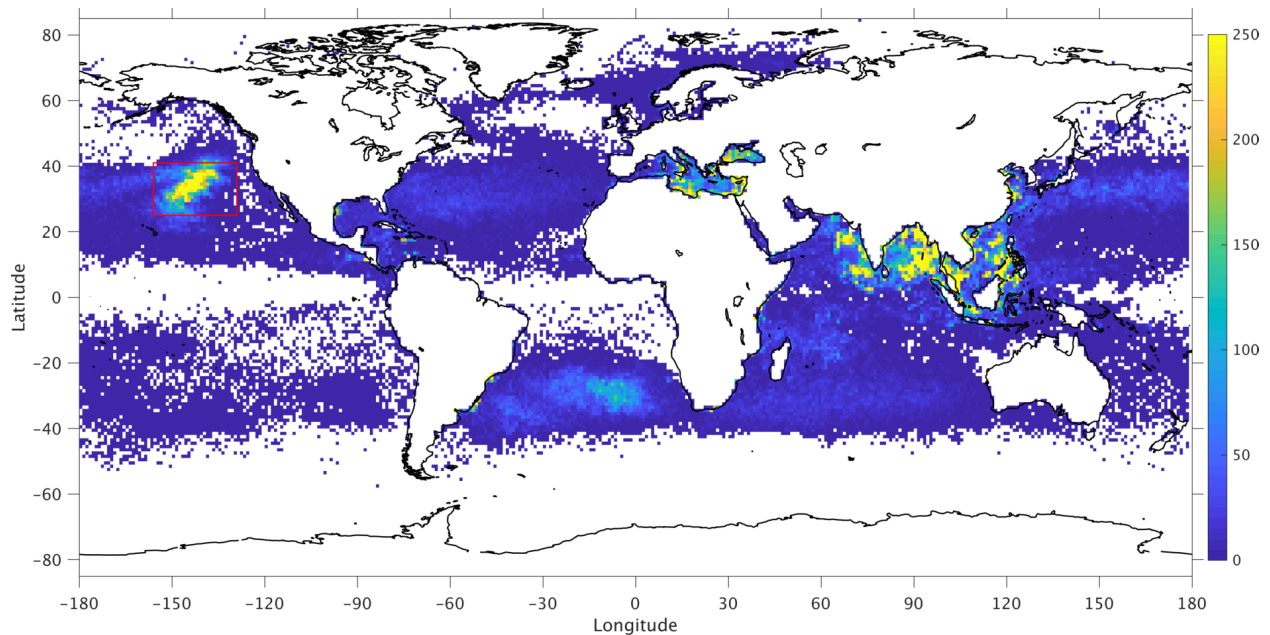


258

259 **Figure 3.** The number of modeled MPW particles in $1^\circ \times 1^\circ$ grid box accumulated in 10 years
 260 (2010-2019) that a) are beached and b) are remain in the water. Out of the total of 3,876,000
 261 released particles, 2,821,752 end up on the beach during the 10-year integration.

262 The question then arises as to whether the amount of the modeled MPW remaining in the ocean
 263 is comparable to the observations. There are very few observations on MPW distribution in the
 264 open oceans and whatever data exists come with large uncertainties. Using data collected across
 265 the World Ocean, Cózar et al. (2014) provided a first-order estimate and found the highest
 266 concentration in the subtropical gyres of the Pacific and Atlantic Oceans in the range of 1-2.5
 267 kg/km^2 . Lebreton et al. (2018) provided a more in-depth estimate of plastics concentration in the
 268 Great Pacific Garbage Patch (GPGP), which is located in the subtropical water between California

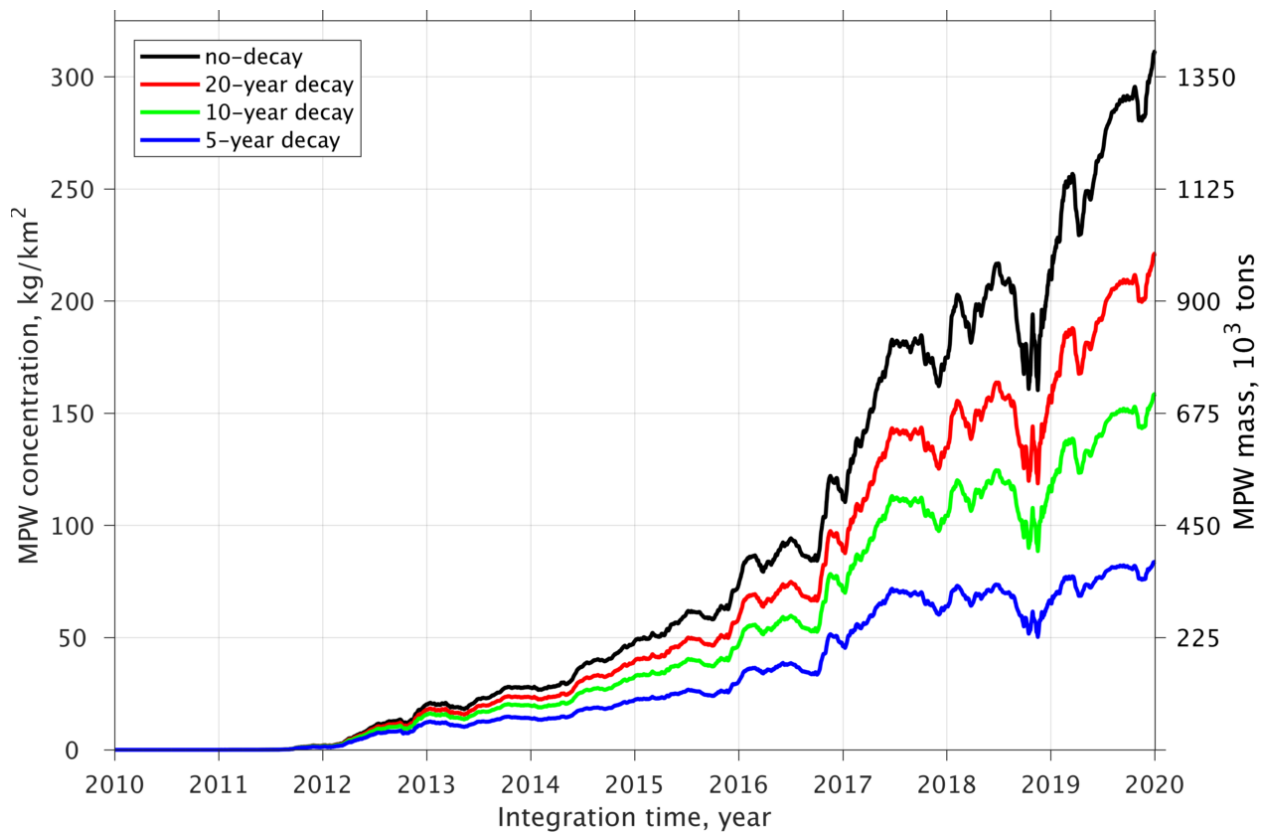
269 and Hawaii, and characterized the MPW into micro (0.05-0.5 cm), meso (0.5-5 cm), macro (5-50
270 cm), and mega (>50 cm) plastics. They estimated that more than 75% of the MPW in the GPGP is
271 in the form of macro and mega plastics larger than 5 cm, whereas micro plastics only accounted
272 for ~8%. The maximum plastic concentrations observed for micro, meso, macro, and mega plastics
273 are 15, 47, 70, and 342 kg/km², respectively, which is one to two orders higher than the highest
274 concentration in Cózar et al. (2014).



275
276 **Figure 4.** The modeled MPW concentration (in kg/km²) showing 10-years of MPW accumulation
277 (2010 to 2019) at the end of the integration. The red box denotes the Great Pacific Garbage Patch
278 (GPGP, 129-156°W, 25-41°N).

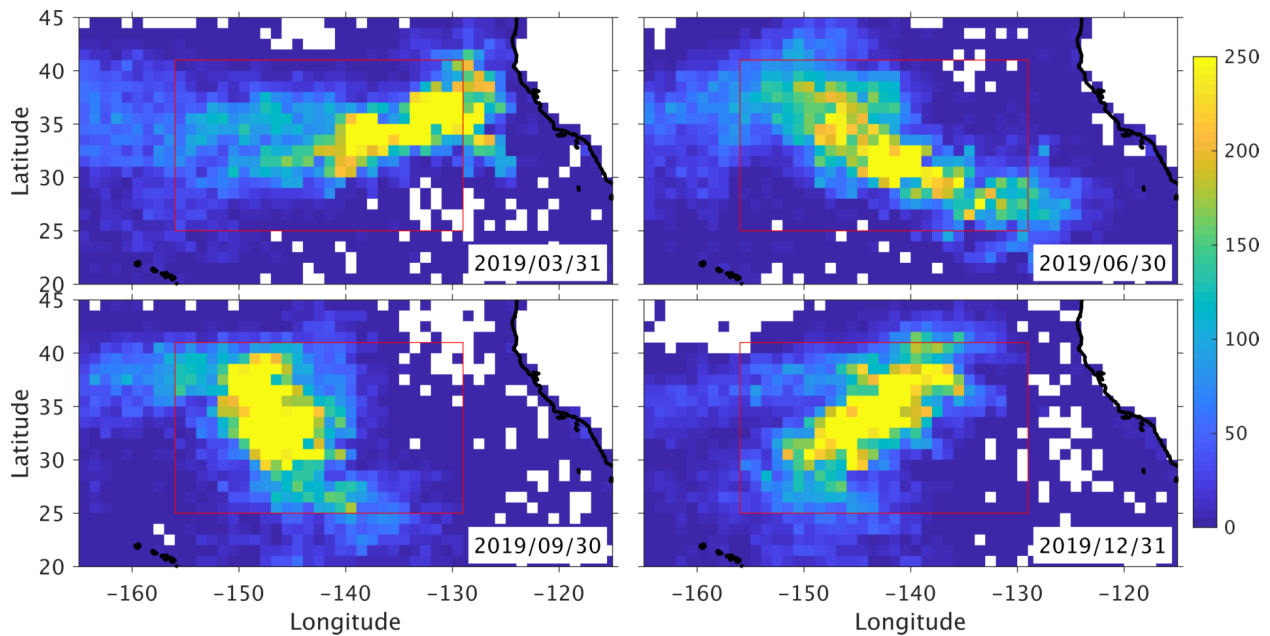
279 Figure 4 displays the distribution of the modeled MPW concentration at the end of the 10-year
280 accumulation. The general pattern (i.e., the high concentrations found in the subtropical regions of
281 the North Pacific and the South Atlantic Oceans, as well as in the Mediterranean Seas) is consistent
282 with Cózar et al. (2014, 2015) and Viatte et al. (2020). High concentrations of the modeled MPW
283 are also found in the northern Indian Ocean and in the marginal seas that connect the Pacific and
284 Indian Oceans. To our knowledge, there are no direct observations in these regions, but this should

285 not come as a surprise given the fact that a majority of the MPW mass that enters the ocean is from
 286 the surrounding South and East Asian countries. Quantitatively, the highest modeled concentration
 287 of MPW in the GPGP is $\sim 500 \text{ kg/km}^2$. This value is of the same order as the highest concentration
 288 reported in Lebreton et al. (2018), but uncertainties are large and our results are strongly dependent
 289 on the decay time scale chosen to represent MPW break down at sea. Figure 5 illustrates the impact
 290 of different decaying time scales on the accumulation of the MPW mass in the GPGP (defined as
 291 $25\text{-}41^\circ\text{N}$, $129\text{-}156^\circ\text{W}$, i.e., red box in Figure 4). In the absence of decay, MPW in the GPGP box
 292 would reach ~ 1.4 million tons after 10 years. With the five-year decay time scale, it reaches a near
 293 steady state of ~ 370 thousand tons or an average of 80 kg/km^2 .



294
 295 **Figure 5.** The accumulation of the modeled MPW concentration or mass in the GPGP area (129-
 296 156°W , $25\text{-}41^\circ\text{N}$, red rectangle box in Figure 4) under different decaying scenarios.

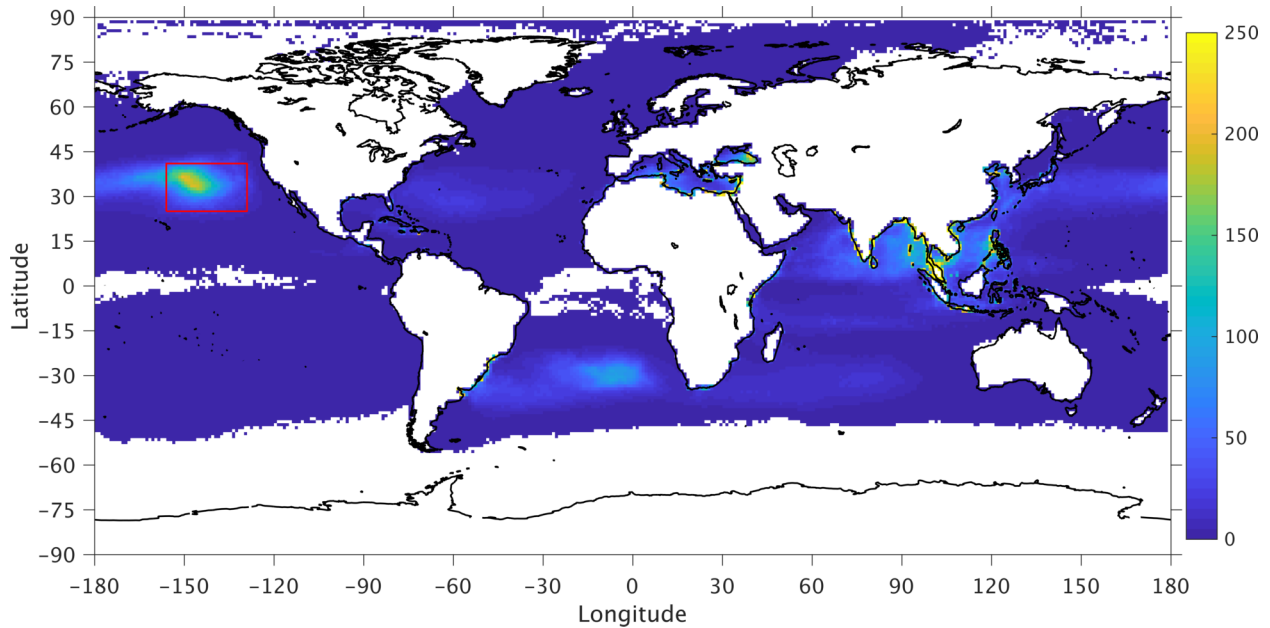
297 Figure 4 represents only a snapshot the MPW concentration at the end of the 10-year
 298 accumulation in 2019 and it is important to note that, as reported by Maes et al. (2016), the
 299 distribution of MPW in the GPCP domain (outlined in red) varies greatly with changes in the ocean
 300 circulation and wind patterns. To illustrate how quickly this distribution can change in time, Figure
 301 6 displays four snapshots of MPW concentration in the GPGP area at the end of March, June,
 302 September, and December 2019, respectively. In March, the GPCP patch is close to the continental
 303 U.S. on the eastern side of the red box while six months later it is close to the western edge of the
 304 red box.



305
 306 **Figure 6.** Snapshots of the modeled MPW concentration (in kg/km^2) near the GPGP area at the
 307 end of March, June, September, and December 2019.

308 Figure 7 displays the time-averaged MPW concentration for the last three years (2017-2019),
 309 when the amount of MPW in the GPGP reached a quasi-steady state (Figure 5). The overall
 310 distribution is quite close to that seen in Figure 4 and agrees well with previously published
 311 modeling studies (e.g., van Sebille et al., 2012; Maes et al., 2018; Lebreton et al., 2018; Viatte et
 312 al., 2020). The averaged mass of MPW in the GPGP box is ~ 300 thousand tons, which is only

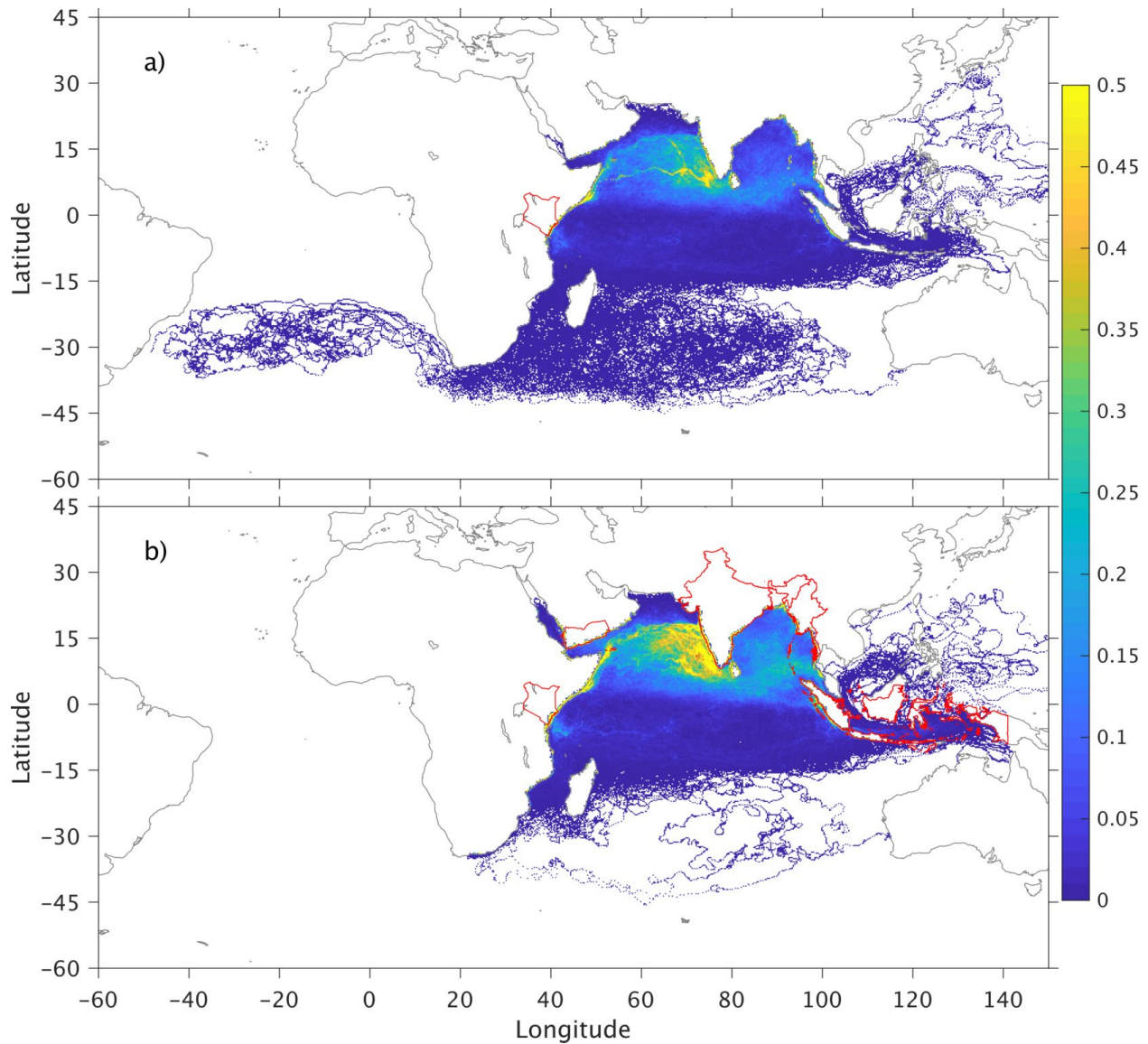
313 slightly lower than the 370 thousand tons present at the end of 2019 (Figure 5). This corresponds
314 to an average concentration of 65 kg/km², which is a little bit higher than the 50 kg/km² estimated
315 by Lebreton et al. (2018).



316
317 **Figure 7.** Similar to Figure 4, but for the time-averaged modeled MPW concentration (in kg/km²)
318 from 2017 to 2019.

319 5.2 Destination and origin of the MPW for an individual country

320 In section 5.1, we showed that modeled MPW concentration in the open ocean is comparable
321 to observational estimates. In this section, we use the model to address the questions raised in the
322 introduction: 1) where does MPW released into the ocean by a given country go and 2) where does
323 MPW found on the coastline of a given country come from. Because observational data were
324 collected on its beaches (Ryan, 2020), which can be used to validate the model, we use Kenya as
325 an example in this section. The statistics for all world countries on MPW destinations and beached
326 MPW sources are provided in the supplement to this article.



327

328 **Figures 8.** Time-averaged concentration (in kg/km²) of the MPW that are released from Kenya
 329 and a) remains in the ocean and b) is beached to another country's coast at the end of the 2019.
 330 Red lines outline the border of Kenya in panel a) and the border of the five countries (Table 1) that
 331 received the most MPW from Kenya (Kenya, India, Indonesia, Myanmar, and Yemen) in panel b).

332

333 The time-averaged concentration of all MPW that originated from Kenya during the 2010-
 334 2019 period is displayed in Figure 8. The figure is divided into two panels, Figure 8a shows the
 335 averaged concentration for those particles that remain at sea at the end of 2019, while Figure 8b
 336 shows the averaged concentration for those particles that ended up on the beach (major countries
 outlined in red) at the end of 2019. The distribution in these two figures is quite similar, which

337 implies that most particles follow a similar pathway: first, they flow northeastward into the Arabian
338 Sea and subsequently into the Bay of Bengal, depending the monsoon currents in the north Indian
339 Ocean, and then eastward following the Equatorial Counter Current (e.g., Tomczak and Godfrey,
340 2003; Shankar et al., 2002). One small difference between the two panels is that some MPW that
341 remains in the ocean is trapped in the subtropical gyre of the South Indian Ocean or escapes into
342 the South Atlantic Ocean via the Agulhas current and associated eddies.

343 Overall, Kenya contributed a total of 58,991 tons of MPW into the ocean for the 2010-2019
344 period, of which 40% (23,596 tons) are considered denser than the seawater and therefore sinks
345 toward the ocean floor near the coast. The remaining 60% or 35,395 tons are considered lighter
346 than seawater and thus are carried by the ocean surface currents and wind. Of the 35,395 tons of
347 lighter than seawater MPW, 20,234 tons (34.3% of the original 58,991 tons) vanishes due to
348 decomposition (the decay process). Table 1 summarizes the destination of the remaining 15,161
349 tons of MPW originating from Kenya: roughly half (52.4%) ends up on the beaches of 26 countries
350 (with only 12 of these countries receiving more than 100 tons) while the rest remains at sea,
351 primarily in the north Indian Ocean. The majority of these recipient countries (outlined in red in
352 Figure 8) are in the northeastern Indian Ocean (India, Indonesia, Myanmar), but some are
353 neighboring countries in Africa (Yemen, Somalia). This distribution is consistent with the surface
354 circulation of the Indian Ocean and slightly over 50% of the MPW originating from Kenya is either
355 back on their beaches (14%) or transported to neighboring countries. By contrast, others countries,
356 such as South Africa or Japan, can have up to 80% of their MPW swept to the ocean interior by
357 strong western boundary currents such as the Agulhas Current or the Kuroshio (see supplement).

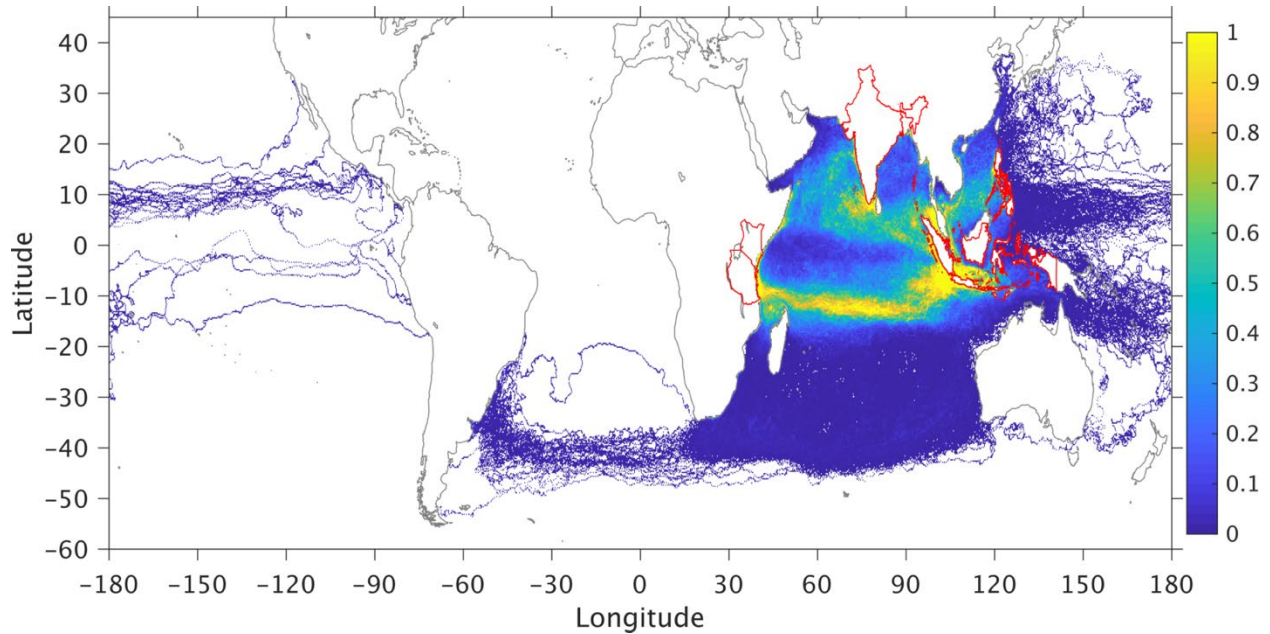
358 **Table 1.** Destination of the modeled MPW that are released from Kenya at the end of the 2010-
 359 2019 accumulation. Note that Kenya is estimated to contribute a total of 58,991 tons MPW into
 360 the ocean in this period and only about ¼ (15,164 tons) is consider here (the rest sinks to the bottom
 361 and/or vanishes due to the decay process).

MPW tons [Percentage]	Destination
7,223 [47.6]	Ocean
7,941 [52.4]	Beached to the following countries*
2,154 [14.2]	Kenya
1,399 [9.2]	India
1,180 [7.8]	Indonesia
703 [4.6]	Myanmar
666 [4.4]	Yemen
441 [2.9]	Somalia
336 [2.2]	Sri Lanka
217 [1.4]	Bangladesh
175 [1.2]	Saudi Arabia
138 [0.9]	Mozambique
130 [0.9]	Thailand
129 [0.9]	Tanzania
91 [0.6]	Oman
53 [0.4]	Madagascar

362
 363 * 12 additional countries that received less than 50 tons MPW (0.4%) are not listed: Sudan,
 364 Malaysia, Djibouti, Eritrea, South Africa, Philippines, Timor-Leste, China, Australia, Vietnam,
 365 Taiwan, and Iran.

366 In the same fashion that MPW from Kenya ends up in other countries, Kenya receives its share
 367 of MPW from other countries (i.e., interconnectivity – see <https://marinelitter.coaps.fsu.edu> and
 368 Appendix for a visualization). Figure 9 displays the time-averaged MPW concentration (in kg/km²)
 369 for the MPW that eventually beach on the Kenyan coast. High-concentrations of MPW are
 370 primarily found near Indonesia and in a latitudinal band around 10°S that spans from Indonesia to
 371 Tanzania that is associated with the westward-flowing South Equatorial Current. While it is not
 372 surprising that most of the MPW that reaches Kenya comes from surrounding countries in the
 373 Indian Ocean, some of the beached MPW can originate from as far as South and Central America

374 (i.e., Brazil, Uruguay, Argentina, Peru, Mexico, Guatemala, and Panama) over a period of less
375 than 10 years.



376
377 **Figure 9.** Time-averaged model MPW concentration (in kg/km²) in the ocean in 2010-2019 for
378 the MPW that are beached on the Kenyan coast by the end of 2019. Red lines outline the border
379 of the five countries that contribute the most MPW to Kenya (Tanzania, Indonesia, India,
380 Philippines, and Kenya; see Table 2).

381 Quantitatively, the model shows that, in ten years, a total of 48,304 tons of MPW from 46
382 countries (Table 2) reached the Kenyan coast (with 19 countries contributing at least 50 tons). The
383 southern neighbor Tanzania contributed the most (38% of the total), which is consistent with MPW
384 being advected by the northward-flowing current along the Tanzanian and Kenyan coasts (Semba
385 et al., 2019). Three southern Asian countries (Indonesian, India, and the Philippines) together
386 contribute 43.5%. These MPW are first carried to the eastern part of the equatorial Indian Ocean,
387 through the Indonesian Throughflow (ITF, Gordon et al., 1997; Metzger et al., 2010) or the
388 Monsoon Current in the North Indian Ocean, before being carried to the Tanzanian/Kenyan coasts
389 via the westward-flowing South Equatorial Current.

390 **Table 2.** Origin of the modeled MPW found along the Kenyan coast, a total of 48,304 tons, along
 391 with the list of country/origin of the bottles that were found during the National Marine Litter Data
 392 Collection Training in August 13-22, 2019 (Ryan, 2020).

MPW tons [percentage]	Origin*	number of bottles**	Origin
18,300 [37.9]	Tanzania	1,227	Kenya
12,640 [26.2]	Indonesia	98	Tanzania
5,209 [10.8]	India	86	Indonesia
3,160 [6.5]	Philippines	42	United Arab Emirates
2,154 [4.5]	Kenya	27	China
1,507 [3.1]	Malaysia	16	India
1,370 [2.8]	Comoros	15	Malaysia
743 [1.5]	Vietnam	14	Mayotte
519 [1.1]	China	13	Madagascar
515 [1.1]	Sri Lanka	9	Comoros
418 [0.9]	Mozambique	8	Thailand
375 [0.8]	Myanmar	5	Vietnam
277 [0.6]	Thailand		
204 [0.4]	Bangladesh		
196 [0.4]	Pakistan		
155 [0.3]	Timor-Leste		
144 [0.3]	South Africa		
115 [0.2]	Somalia		
51 [0.1]	Madagascar		

393 * 27 countries contributed less than 50 tons MPW (0.1%) are not listed: Yemen, Taiwan, Solomon
 394 Islands, Brazil, Papua New Guinea, Uruguay, Australia, Djibouti, Peru, Argentina, Mexico,
 395 Maldives, Vanuatu, Brunei, Guatemala, Oman, Seychelles, Fiji, Mauritius, Japan, El Salvador,
 396 Panama, Iran, Eritrea, Samoa, Palau, and Micronesia.

397 ** (many) countries with less than five bottles are not listed

398 Because in-situ data are difficult to collect, not much exists that can be used to validate the model.

399 It is already challenging to quantify the amount of plastics found on beaches (often remote), let
 400 alone to provide further reliable information on its origin. The reason Kenya was chosen as the
 401 example here is that data collected in Kenya during the National Marine Litter Data Collection
 402 Training in August 13-22, 2019 (Ryan, 2020) are available, which can provide some perspective
 403 of the model results. That data collection team found a total of 1,819 plastic bottles on Kenyan
 404 beaches during the 10-day training period with about two-thirds (1,227) determined to be of local
 405 origin and from Kenya. For those identified as coming from outside of Kenya, the two countries

406 that contribute the most ocean MPW to Kenya in our global model (Tanzania and Indonesia) are
407 also the top two countries in the bottle counts (Table 2). Other countries, such as China, India, and
408 Malaysia, are also among the key contributors in both the (bottle counts) data and the modeled
409 MPW. Clearly, such a comparison is limited, but a reasonable agreement exists between the model
410 and the observations. Overall, MPW found on the Kenyan coast has two major origins: 1) East
411 African countries, its southern neighbor Tanzania in particular, and 2) South Asian countries (and
412 the islands in the western Indian Ocean on the path of the South Equatorial Current).

413 **6. Summary and discussion**

414 In summary, using worldwide estimates of MPW provided by Lebreton et al (2017) and
415 Lebreton and Andrady (2019), we are able to provide a quantitative global estimate of 1) where
416 does MPW released into the ocean by a given country go and 2) where does MPW found on the
417 coastline of a given country come from. Tables summarizing the statistics for all world countries
418 can be accessed from the supplemental information in .pdf or .csv formats. The overall distribution
419 of the modeled MPW is in good agreement with the limited observations that we have at our
420 disposal and with previous studies. However, observations of MPW that can be used to validate
421 the model are extremely scarce and it is difficult, not only to quantify the amount of plastic found
422 on beaches (often remote), but also to have any information on its origin. As shown in section 5,
423 the numerical results are consistent with data collected in Kenya during the National Marine Litter
424 Data Collection Training (Ryan et al., 2020), but we would need many more measurements of this
425 kind from many countries to have a more accurate estimate of the origin of MPW found on the
426 coastline. This is further complicated by the fact that a lot of MPW released in the ocean by a
427 country do not necessarily originate from that country. Law et al. (2020) estimates that more than
428 half of all plastics collected for recycling in the U.S. are shipped abroad and that 88% (~1 million

429 metric tons) of the exported plastic went to countries that struggle to effectively manage, recycle,
430 or dispose of plastics.

431 This modeling study has limitations in that it does not fully take into account the life cycle of
432 the plastic at sea (approximated using a five-year decay time scale), nor does it take into account
433 the size of the litter (macro to nano) and differences in windage. There are also uncertainties
434 associated with ocean currents and the winds used to move the MPW in the ocean. Nonetheless, it
435 does provide first-order numbers that can be used by governments, non-profit organizations, and
436 the general public to redirect or reinforce actions to reduce the amount of marine litter. This is
437 especially important since a recent publication by Borrelle et al. (2020) estimates that in the next
438 10 years the plastic waste that enters into waterways and ultimately the oceans could reach 22
439 million tons and possibly as much as 58 million tons a year. And, this estimation takes into account
440 the thousands of commitments made by the government and the industry to reduce plastic
441 pollution.

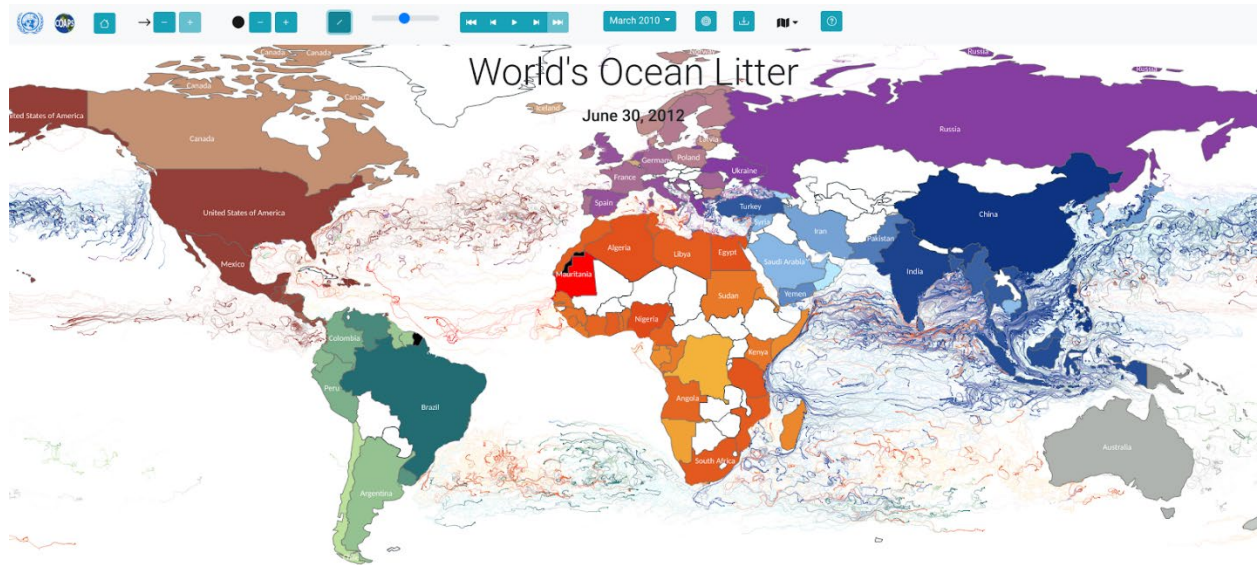
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446 the paper.

447 **Appendix: Web interface**

448 A user-friendly website was developed (<https://marinelitter.coaps.fsu.edu/>) to present the
449 model results in a dynamic and efficient manner. Twelve five-year monthly releases of MPW are
450 displayed with dynamic animations of streamlines that show the marine litter path through time.

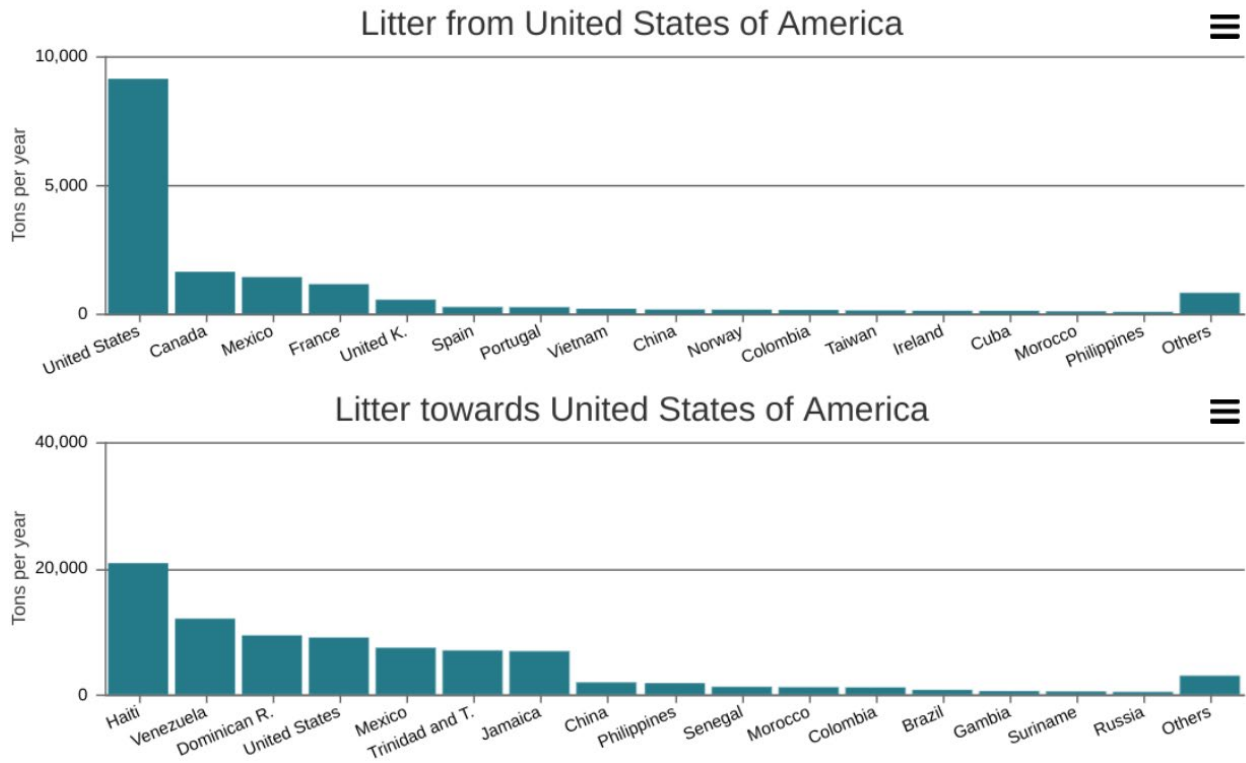
451 The color palettes in the map vary by continent, and the specific color for each country is
452 proportional to the total amount of litter generated. Figure 10 shows a screenshot of this interface
453 and the colors assigned to the continents and their corresponding countries.



454
455 **Figure 10.** Example of the web interface generated to display global litter paths per country.

456 The web interface also provides information about individual statistics for each country, which
457 includes the tons of litter generated each year, the percentage of marine litter that stay in the oceans,
458 and the amount of litter that ends up on the beach. , explicitly describing the amount of litter that
459 ends on each country. Each country's statistics are provided as bar plots, and the raw data can be
460 downloaded from the website in several file formats (.pdf, .csv, and .json). Figure 11 shows an
461 example of the ocean litter statistics for the United States.

United States of America exports 54,233 tons in ten years
 37,323 (68.8%) end up in the ocean and 16,910 (31.2%) end up on the beach.



462

463 **Figure 11.** Ocean litter statistics for the United States, as shown on the web interface.

464 For an efficient display of the marine litter paths, only a subset of the total simulated particles
 465 is shown for each monthly release (half for desktop applications and one fourth for mobile
 466 browsers), accounting for up to 14 million particle locations in each five-year animation. Finally,
 467 the interface empowers the user with multiple animation and cosmetic controls to quickly identify
 468 the marine debris pathways through time.

469 **References**

- 470 Arduin, F., B. Chapron, and F. Collard, 2009. Observation of swell dissipation across oceans.
 471 *Geophys. Res. Lett.*, **36**, L06607, doi:10.1029/2008GL037030.
- 472 Andrady, A. L., 2011. Microplastics in the marine environment. *Mar. Pollut. Bull.*, **62**(8), 1596–
 473 1605. doi:10.1016/j.marpolbul.2011.05.030
- 474 Barrett J., Z. Chase, J. Zhang, M. M. B. Holl, K. Willis, A. Williams, B. D. Hardesty, and C.
 475 Wilcox, 2020. Microplastic Pollution in deep-sea sediments from the Great Australian Bight.
 476 *Front. Mar. Sci.*, **7**, 576170, doi:10.3389/fmars.2020.576170
- 477 Bergmann, M., M. Tekman, and L. Gutow, 2017. Sea change for plastic pollution. *Nature*, 544,
 478 297, doi:10.1038/544297a
- 479 Borrelle, S.B., J. Ringma, K.L. Law, C.C. Monnahan, L. Lebreton, A. McGivern, E. Murphy, J.
 480 Jambeck, G.H. Leonard, M.A. Hilleary, M. Eriksen, H.P. Possingham, H. De Frond, L.R.
 481 Gerber, B. Polidoro, A. Tahir, M. Bernard, N. Mallos, M. Barnes, and C.M. Rochman, 2020.
 482 Predicted growth in plastic waste exceeds efforts to mitigate plastic pollution. *Science*, 1515–
 483 1518. doi:10.1126/science.aba3656
- 484 Breivik, Ø., A. A. Allen, C. Maisondieu, and J. C. Roth, 2011. Wind-induced drift of objects at
 485 sea: the leeway field method. *Appl. Ocean Res.*, **33**, 100–109, doi:10.1016/j.apor.2011.01.005.
- 486 Chassignet, E. P., H. E. Hurlburt, E. J. Metzger, O. M. Smedstad, J. Cummings, G. R. Halliwell,
 487 R. Bleck, R. Baraille, A. J. Wallcraft, C. Lozano, H. L. Tolman, A. Srinivasan, S. Hankin, P.
 488 Cornillon, R. Weisberg, A. Barth, R. He, F. Werner, and J. Wilkin, 2009. U.S. GODAE: Global
 489 Ocean Prediction with the HYbrid Coordinate Ocean Model (HYCOM). *Oceanography*, **22**(2),
 490 64-75. doi: 10.5670/oceanog.2009.39
- 491 Chubarenko I., A. Bagaev, M. Zobkov, and E. Esiukova, 2016. On some physical and dynamical
 492 properties of microplastic particles in marine environment. *Mar. Pollut. Bull.*, **108**, 105–112,
 493 doi:10.1016/j.marpolbul.2016.04.048.
- 494 Cózar A., F. Echevarría, J.I. González-Gordillo, X. Irigoien, B. Úbeda, S. Hernández-León, Á. T.
 495 Palma, S. Navarro, J. García-de-Lomas, A. Ruiz, and M.L. Fernández-de-Puelles, 2014. Plastic
 496 debris in the open ocean. *Proceedings of the National Academy of Sciences*, **111**(28), 10239–
 497 10244, doi:10.1073/pnas.1314705111
- 498 Cózar A., M. Sanz-Martín, E. Martí, J.I. González-Gordillo, B. Ubeda, J.Á. Gálvez, X. Irigoien,
 499 and C.M. Duarte, 2015. Plastic accumulation in the Mediterranean Sea. *PLoS ONE*, **10**(4),
 500 e0121762, doi:10.1371/journal.pone.0121762
- 501 Cummings J. A. and O. M. Smedstad. 2013: Variational Data Assimilation for the Global Ocean.
 502 Data Assimilation for Atmospheric, *Oceanic and Hydrologic Applications vol II*, chapter 13,
 503 303-343. doi:10.1007/978-3-642-35088-7_13
- 504 Delandmeter, P., and E. van Sebille, 2019: The Parcels v2.0 Lagrangian framework: new field
 505 interpolation schemes, *Geoscientific Model Development*, **12**, 3571-3584. doi:10.5194/gmd-
 506 12-3571-2019
- 507 Geyer, R., J. Jambeck, and K. L. Law, 2017. Production, use, and fate of all plastics ever made.
 508 *Sci. Adv.*, **3**(7):e1700782.
- 509 Gordon, A. L., S. Ma, D. B. Olson, P. Hacker, A. Ffield, L. D. Talley, D. Wilson, and M. Baringer,
 510 1997. Advection and diffusion of Indonesian through flow water within the Indian Ocean
 511 South Equatorial Current. *Geophys. Res. Lett.* **24**, 2573-2576.

512 Hardesty, B. D., T. J. Lawson, T. van der Velde, M. Lansdell, and C. Wilcox, 2017. Estimating
513 quantities and sources of marine debris at a continental scale. *Frontiers in Ecology and the*
514 *Environment*, **15**(1), 18-25, doi:10.1002/fee.1447

515 Helber, R. W., T. L. Townsend, C. N. Barron, J. M. Dastugue and M. R. Carnes, 2013. Validation
516 Test Report for the Improved Synthetic Ocean Profile (ISOP) System, Part I: Synthetic Profile
517 Methods and Algorithm. NRL Memo. Report, NRL/MR/7320—13-9364.

518 Hoornweg, D., and P. Bhada-Tata, 2012. What a waste: a global review of solid waste
519 management. *Urban development series; knowledge papers no. 15*. World Bank.
520 <http://hdl.handle.net/10986/17388>.

521 Jambeck, J. R., R. Geyer, C. Wilcox, T. R. Siegler, M. Perryman, A. L. Andrady, R. Narayan, and
522 K.L. Law, 2015. Plastic waste inputs from land into the ocean, *Science*, **347**, 768–771.
523 doi:10.1126/science.1260352

524 Kimukai, H., K. Yoichi, K. Koushirou, O. Masaki, Y. Kazunori, H. Toshihiko, and S. Katsuhiko,
525 2020. Low temperature decomposition of polystyrene. *Appl. Sci.*, **10**, 15:5100,
526 doi:10.3390/app10155100

527 Kinsman, B., 1965. *Wind Waves*, Prentice-Hall, 660pp

528 Kubota M., 1994. A mechanism for the accumulation of floating marine debris north of Hawaii. *J*
529 *Phys. Oceanogr.* **24**, 1059–1064.

530 Kukulka T., G. Proskurowski, S. Morét-Ferguson, D. W. Meyer, and K. L. Law. 2012. The effect
531 of wind mixing on the vertical distribution of buoyant plastic debris. *Geophys. Res. Lett.*, **39**,
532 L07601, doi:10.1029/2012GL051116.

533 Law, K. L., N. Starr, T. R. Siegler, J. R. Jambeck, N. J. Mallos and G. H. Leonard, 2020. The
534 United States’ contribution of plastic waste to land and ocean. *Science*
535 *Advances*, doi:10.1126/sciadv.abd0288.

536 Lebreton, L., J. van der Zwet, J.W. Damsteeg, B. Slat, A. Andrady, and J. Reisser, 2017. River
537 plastic emissions to the world’s oceans. *Nat. Commun.*, **8**, 15611, doi:10.1038/ncomms15611.

538 Lebreton, L., B. Slat, F. Ferrari, B. Sainte-Rose, J. Aitken, R. Marthouse, S. Hajbane, S. Cunsolo,
539 A. Schwarz, A. Levivier, and K. Noble, 2018. Evidence that the Great Pacific Garbage Patch
540 is rapidly accumulating plastic. *Sci. Rep.*, **8**, 4666, doi :10.1038/s41598-018-22939-w

541 Lebreton, L., and Andrady, A., 2019. Future scenarios of global plastic waste generation and
542 disposal. *Palgrave Commun* **5**, 6, doi:10.1057/s41599-018-0212-7.

543 Maes C., B. Blanke, and E. Martinez, 2016. Origin and fate of surface drift in the oceanic
544 convergence zones of the eastern Pacific. *Geophys. Res. Lett.*, **43**, 3398–3405,
545 doi:10.1002/2016G L0682 17.

546 Maes C., N. Grima, B. Blanke, E. Martinez, T. Paviet-Salomon, and T. Huck, 2018. A surface
547 “superconvergence” pathway connecting the South Indian Ocean to the subtropical South
548 Pacific gyre. *Geophys. Res. Lett.*, **45**, 1915–1922, doi:10.1002/2017GL076366.

549 Metzger, E. J., H. E. Hurlburt, X. Xu, A. L. Gordon, J. Sprintall, R. D. Susanto, and H. M. van
550 Aken, 2010. Simulated and observed circulation in the Indonesian Seas: 1/12° global HYCOM
551 and the INSTANT observations, *Dynamics of Atmospheres and Oceans*,
552 doi:10.1016/j.dynatmoce. 2010.04.002

553 Metzger, E. J., O. M. Smedstad, P. G. Thoppil, H. E. Hurlburt, J. A. Cummings, A. J. Wallcraft,
554 L. Zamudio, D. S. Franklin, P. G. Posey, M. W. Phelps, P. J. Hogan, F. L. Bub, and C. J.
555 DeHaan, 2014: US Navy operational global ocean and Arctic ice prediction systems.
556 *Oceanography*, **27**(3), 32-43, doi:10.5670/oceanog.2014.66.

557 Metzger, E., R.W. Helber, P.J. Hogan, P.G. Posey, P.G. Thoppil, T.L. Townsend, A.J. Wallcraft,
558 O.M. Smedstad, D.S. Franklin, L. Zamudo, and M.W. Phelps, 2017. Global Ocean Forecast
559 System 3.1 validation test. *Naval Research Laboratory, Stennis Space Center MS*, Tech. rep.
560 NRL/MR/7320--17-9722, 61pp.

561 Moore, C. J., S. L. Moore, M. K. Leecaster, and S. B. Weisberg, 2001. A comparison of plastic
562 and plankton in the North Pacific Central Gyre, *Marine Pollution Bulletin*, 42, 12, 1297-1300,
563 doi:10.1016/S0025-326X(01)00114-X.

564 Pereiro, P., C. Carlos Souto, and J. Gago, 2018. Calibration of a marine floating litter transport
565 model, *J. Operational Oceanogr.*, 11:2, 125-133, doi:10.1080/1755876X.2018.1470892.

566 Plastics Europe, 2017. [https://www.plasticseurope.org/en/resources/publications/274-plastics-](https://www.plasticseurope.org/en/resources/publications/274-plastics-facts-2017)
567 [facts-2017](https://www.plasticseurope.org/en/resources/publications/274-plastics-facts-2017)

568 Ryan, P. G., 2020. Land or sea? What bottles tell us about the origins of beach litter in Kenya,
569 *Waste Management*, 116, 49–57, doi:10.1016/j.wasman.2020.07.044

570 Schmidt, C., T. Krauth, and S. Wagner, 2017. Export of plastic debris by rivers into the sea.
571 *Environ. Sci. Technol.*, 51(21), 12246–12253. doi:10.1021/acs.est.7b02368

572 Semba, M., R. Lumpkin, I. Kimirei, Y. Shaghude, and N. Nyandwi, 2019. Seasonal and spatial
573 variation of surface current in the Pemba Channel, Tanzania. *PLoS ONE*, 14(1), e0210303,
574 doi:10.1371/journal.pone.0210303

575 Shankar, D., P.N. Vinayashandran, and A. S. Unnikrishnan, 2002. The monsoon currents in the
576 north Indian Ocean, *Progress in Oceanography*, 52, 63-120.

577 Special issue on nanoplastic, 2019: Nanoplastic should be better understood. *Nat. Nanotechnol.*,
578 14, 299, doi:10.1038/s41565-019-0437-7

579 Tomczak, M., and J. S. Godfrey, 2003. Regional Oceanography: an Introduction. 2nd Edition.

580 Thompson, R. C., Y. Olsen, R. P. Mitchell, A. Davis, S. J. Rowland, A. W. G. John, D. McGonigle,
581 and A. Russell, 2004. Lost at sea: Where is all the plastic? *Science*, 304, 838–838,
582 doi:10.1126/science.1094559.

583 van den Bremer, T. S., and Ø. Breivik, 2018. Stokes drift. *Phil. Trans. R. Soc. A.*, 37620170104,
584 doi:10.1098/rsta.2017.0104.

585 van Sebille E., M. H. England, and G. Froyland, 2012. Origin, dynamics and evolution of ocean
586 garbage patches from observed surface drifters. *Environ. Res. Lett.*, 7, 044040.

587 van Sebille, E., S. Aliani, K. L. Law, N. Maximenko, J. M. Alsina, A. Bagaev, M. Bergmann, B.
588 Chapron, I. Chubarenko, A. Cózar, P. Delandmeter, M. Egger, B. Fox-Kemper, S. P. Garaba,
589 L. Goddijn-Murphy, B. D. Hardesty, M. J. Hoffman, A. Isobe, C. E. Jongedijk, M. L. A.
590 Kaandorp, L. Khatmullina, A. A. Koelmans, T. Kukulka, C. Laufkötter, L. Lebreton, D.
591 Lobelle, C. Maes, V. Martinez-Vicente, M.A. Morales Maqueda, M. Poulain-Zarcos, E.
592 Rodríguez, P. G. Ryan, A. L. Shanks, W. J. Shim, G. Suaria, M. Thiel, T. S. van den Bremer,
593 and D. Wichmann, 2020. The physical oceanography of the transport of floating marine debris.
594 *Env. Res. Lett.*, 15, 023003. doi:10.1088/1748-9326/ab6d7d

595 Viatte, C., C. Clerbaux, C. Maes, P. Daniel, R. Garello, S. Safieddine, and F. Arduin, 2020. Air
596 pollution and sea pollution seen from space. *Surveys in Geophysics*, doi:10.1007/s10712-020-
597 09599-0

598 Woodall, L.C., A. Sanchez-Vidal, M. Canals, G.L.J. Paterson, R. Coppock, V. Sleight, A. Calafat,
599 A.D. Rogers, B.E. Narayanaswamy, and R.C. Thompson, 2014. The deep sea is a major sink
600 for microplastic debris. *R. Soc. Open Sci.*, 1:8, doi:10.1098/rsos.140317